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- 1 Authors: James R. Collins^{1,2*†}, Paul D. Fucile³, Glenn McDonald⁴, Justin E. Ossolinski², Richard G.
- 2 Keil⁵, James R. Valdes³, Scott C. Doney², and Benjamin A. S. Van Mooy^{2*}
- 3 **Title:** An autonomous, *in situ* light-dark bottle device for determining community respiration and net
- 4 community production
- 5 Affiliations: ¹ MIT/WHOI Joint Program in Oceanography/Applied Ocean Science and Engineering,
- 6 Woods Hole, Massachusetts, USA
- 7 ² Department of Marine Chemistry and Geochemistry, Woods Hole Oceanographic Institution,
- 8 Woods Hole, Massachusetts, USA
- 9 ³ Department of Physical Oceanography, Woods Hole Oceanographic Institution, Woods Hole, MA
- 10 02543, USA
- ⁴ Department of Applied Ocean Physics and Engineering, Woods Hole Oceanographic Institution,
- 12 Woods Hole, MA 02543, USA
- ⁵ School of Oceanography, University of Washington, Seattle, WA 98195, USA
- 14 * Correspondence: James R. Collins, james.r.collins@aya.yale.edu or Benjamin A. S. Van Mooy,
- 15 bvanmooy@whoi.edu
- 16 *†* **Present Address:** James R. Collins, School of Oceanography and eScience Institute, University of
- 17 Washington, Seattle, WA 98195, USA
- 18 Running head: Autonomous observations of community metabolism
- 19 Keywords: respiration, community metabolism, aquatic microbial ecology, autonomous
- 20 instrumentation, optodes, dissolved oxygen, ocean observing

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Citation: Collins, J. R., P. D. Fucile, G. McDonald, J. E. Ossolinski, R. G. Keil, J. R. Valdes, S. C. Doney, and B. A. S. Van Mooy. 2018. An autonomous, *in situ* light-dark bottle device for determining community respiration and net community production. *Limnology & Oceanography: Methods*; doi:<u>10.1002/lom3.10247</u>

22 Abstract

23 We describe a new, autonomous, incubation-based instrument that is deployed *in situ* to 24 determine rates of gross community respiration and net community production in marine and aquatic 25 ecosystems. During deployments at a coastal pier and in the open ocean, the PHORCYS 26 (PHOtosynthesis and Respiration Comparison-Yielding System) captured dissolved oxygen fluxes 27 over hourly timescales that were missed by traditional methods. The instrument uses fluorescence-28 quenching optodes fitted into separate light and dark chambers; these are opened and closed with 29 piston-like actuators, allowing the instrument to make multiple, independent rate estimates in the 30 course of each deployment. Consistent with other studies in which methods purporting to measure 31 the same metabolic processes have yielded divergent results, respiration rate estimates from the 32 PHORCYS were systematically higher than those calculated for the same waters using a traditional 33 two-point Winkler titration technique. However, PHORCYS estimates of gross respiration agreed 34 generally with separate incubations in bottles fitted with optode sensor spots. An Appendix describes 35 a new method for estimating uncertainties in metabolic rates calculated from continuous dissolved 36 oxygen data. Multiple successful, unattended deployments of the PHORCYS represent a small step 37 toward fully autonomous observations of community metabolism. Yet the persistence of unexplained 38 disagreements among aquatic metabolic rate estimates — such as those we observed between rates 39 calculated with the PHORCYS and two existing, widely-accepted bottle-based methods - suggests 40 that a new community intercalibration effort is warranted to address lingering sources of error in 41 these critical measurements.

42

44 Introduction

45 Accurate, reproducible, and cost-effective estimates of aerobic respiration and primary 46 production in aquatic systems are essential for research across a diverse array of disciplines in the 47 environmental sciences (del Giorgio and Williams, 2005; Volkmar and Dahlgren, 2006; Staehr et al., 48 2012). Rate measurements of these two metabolic parameters can be applied to various problems, 49 including validating the biogeochemical components of global climate models (Denman et al., 2007), 50 determining the trophic status of surface-water planktonic communities in open ocean ecosystems 51 (Williams, 1998), measuring rates of biological oxygen demand (BOD) in treated wastewater 52 (Spanjers et al., 1994), and identifying unexpected metabolisms in the deep ocean (Reinthaler et al., 53 2010).

54 While an increasing demand for metabolic rate data has encouraged the development of many 55 different methods for estimating rates of photosynthesis (Ducklow and Doney, 2013), the number of 56 new methods for measuring aerobic respiration at the community scale has lagged behind 57 considerably (del Giorgio and Williams, 2005). The majority of field-based methods for measuring 58 rates of respiration and primary production in the ocean fall largely into two categories: (1) in situ 59 geochemical tracer techniques that track changes in the concentration or isotopic composition of 60 dissolved oxygen and carbon dioxide within ocean water masses (e.g., the surface mixed layer), and 61 (2) *in vitro* incubation techniques that track the rates at which plankton exchange oxygen or carbon 62 dioxide in discrete seawater samples (i.e., bottle incubations). The merits and faults of these two 63 categories of approaches have been vigorously debated while significant and often unexplained 64 differences are noted in the rate estimates they yield (Duarte et al., 2013; Ducklow and Doney, 2013; 65 Williams et al., 2013). The former category has benefited considerably from recent advances in 66 optical sensor technology (Moore et al., 2009), mass spectrometry (Goldman et al., 2015), and 67 techniques for analysis of optical sensor data from autonomous underwater vehicles (Nicholson et al., 2015). By maximizing the extent to which sensors are integrated into the surrounding environment,
low-power instruments increase the spatial and temporal resolution of geochemical tracers *in situ* and
permit increasingly autonomous, long-term deployments (Prien, 2007; Riser and Johnson, 2008;
Porter et al., 2009).

72 By contrast, the field has seen relatively few technical advances in *in vitro* incubation 73 techniques. In vitro techniques provide an important complement to in situ methods because they are 74 sensitive to short-term perturbations and are amenable to experimental design. For these reasons, the traditional two-point light and dark bottle incubation technique (Gaarder and Gran, 1927) and the ¹⁴C 75 76 incubation method (Steeman Nielsen, 1952) continue to dominate incubation-based studies, although 77 a number of other methods based on electron transport (e.g., Kenner and Ahmed, 1975) or fluxes of ¹⁸O or CO₂ (Bender et al., 1987; Robinson and Williams, 2005) have been introduced over the course 78 79 of the last half-century. A number of these methods have been incorporated into modern designs for 80 benthic flux chambers and so-called "benthic landers," enabling investigators to capture fluxes of 81 oxygen and other gases *in situ* at the sediment-water interface instead of in core samples aboard ship 82 (Hammond et al., 2004; compare, e.g., Martens et al., 2016, Fuchsman et al., 2015, or Lee et al., 83 2015, to Kim et al., 2015). However, even the most advanced of these devices can require the use of 84 divers or remotely operated vehicles (ROVs) for deployment, maintenance, or recovery. 85 Additionally, by their nature, few of these designs can be programmed to conduct multiple 86 incubations over the course of a single deployment. Taylor et al. addressed this obstacle with the 87 submersible incubation device (SID), which for the first time allowed multiple, unattended incubations with ¹⁴C-bicarbonate to be conducted *in situ* (Taylor and Doherty, 1990; Taylor et al., 88 89 1993). The SID represented a significant advance but was limited by its relatively small 400 mL 90 incubation chamber and its reliance on the use of radiolabeled reagents.

91 Among the advances most consequential for *in situ* instrumentation was the adaptation to 92 marine applications of optical technologies such as optodes (e.g., Klimant et al., 1995; Tengberg et 93 al., 2006) and optode sensor spots (e.g., Warkentin et al., 2007), which exploit fluorescence 94 (luminescence) quenching to measure dissolved oxygen concentrations non-destructively and without 95 themselves contributing to oxygen consumption in the sample. Integral optodes and sensor spots 96 based on the same technology have now been successfully used in a variety of shipboard 97 configurations to measure rates of gross community respiration in whole, unconcentrated and 98 unfiltered water samples and in water containing particle material from marine and aquatic 99 environments (Edwards et al., 2011; Wikner et al., 2013; Collins et al., 2015). More recently, 100 shipboard in vitro measurements of respiration within individual marine particles were made 101 successfully using oxygen microelectrodes (Belcher et al., 2016a; Belcher et al., 2016b). A 102 significant recent advance was also achieved with the RESPIRE device, which uses an optode fitted 103 into a modified sediment trap to make particle respiration measurements in situ (Boyd et al., 2015; 104 McDonnell et al., 2015).

105 Despite the significant progress represented in these optode-driven systems, incubation-based 106 methods remain prone to a number of sources of error that demand reconciliation. These can be 107 generally divided into two categories: (1) those that result from the preparation for or act of 108 incubating natural microbial populations and (2) errors inherent in the method used to determine the 109 concentration of dissolved oxygen (e.g., Winkler titration, fluorescence-quenching optode, or Clark 110 electrode). The sources of uncertainty associated with bottle/chamber incubations span both 111 categories and include (1) contamination, disruption, or bias introduced through the process of 112 obtaining seawater samples from depth and preparing them for incubation (Suter et al., 2016; 113 Tamburini et al., 2013); (2) unrepresentative incubation conditions that do not faithfully reproduce 114 the variations in temperature, turbulence, and light inherent in natural systems; (3) so-called "bottle

115 effects" associated with low-volume incubations which may limit nutrient availability (Furnas, 2002) 116 or induce unnatural changes in community structure (Venrick et al., 1977; Calvo-Díaz et al., 2011); 117 and (4), in the case of metabolic rate measurements extrapolated from Winker (1888) titrations, the 118 lack of temporal resolution inherent in measurements based only on two endpoints. In any study 119 where incubations are used, the choice of incubation methodology places inherent limits on the 120 spatial and temporal resolution of the data collected (Karl et al., 2001). The integration of point 121 measurements — data sparse in time and/or space, whether based on in situ observations or 122 incubations — also creates significant representation error. One solution to this problem is to greatly 123 increase the number of measurements made during data collection using automated technologies. 124 We describe here the *PHO*tosynthesis and *Respiration Comparison-Yielding System* 125 (PHORCYS), a large-volume (i.e., > 2.5 L), light and dark chamber incubation system for 126 autonomous measurement of rates of primary production and respiration at high temporal resolution 127 and under *in situ* conditions. In designing the instrument, we endeavored to minimize the major 128 hypothesized sources of uncertainty associated with traditional incubation-based methods while 129 constructing a system that functions autonomously and interrogates water samples non-destructively. 130 We also sought to eliminate or reduce the need for repeated wet-chemical field measurements such as 131 Winkler oxygen titrations or reagent-based methods used in other autonomous systems. We first 132 describe design and validation of the PHORCYS using two independent methods, and then present 133 results of several deployments of the instrument in different ecosystem types.

134 Materials and procedures

135 Instrument design and operation

The PHORCYS is composed of only a few basic components, making the design highly
scalable and cost-effective (Fig. 1a,b). In nearly all instances, "off the shelf" components of different

138 size or capacity can be easily substituted for those we describe here. The PHORCYS consists of two 139 polycarbonate plastic chambers (usable vol. 5.7 L; Table 1), several auxiliary sensors for collection 140 of environmental data in the ambient water mass outside the chambers, and a watertight power 141 supply, control, and data recording module (Fig. 1a). A piston-like, magnetically coupled actuator is 142 programmed to open and close each chamber at a specified interval, allowing users to perform 143 multiple, unattended incubations over the course of a single deployment. The chamber seals are 144 tapered to avoid the use of rubber O-rings that might have introduced a source of organic 145 contamination into the sample water. In experiments, the polycarbonate plastic used for the 146 PHORCYS incubation chambers reduced photosynthetically active radiation (PAR) within the 147 transparent chamber to 83% of incident strength; we present a means of accounting for this 148 attenuation, below. The opaque chamber was darkened by application of a coating to the outside of 149 the cylinder. All PHORCYS components are mounted to a stainless-steel frame, allowing the 150 instrument to operate to a depth of 100 m.

151 Dissolved oxygen concentrations in the two chambers are monitored with fast-response. 152 fluorescence quenching oxygen optodes (Aanderaa model 4531D; accuracy $< 8 \mu M O_2$; resolution <153 1 μ M O₂; response time < 30 s; Aanderaa Data Instruments, Inc., Bergen, Norway); each optode is 154 fitted into its chamber using a water- and gas-tight flange assembly. The instrument also has several 155 external sensors, including a third optode to monitor dissolved oxygen concentrations in the water 156 mass outside of the two chambers, PAR sensor, beam transmissometer, chlorophyll fluorometer, and 157 CTD. We did not investigate whether the arrangement of these sensors had any effect on the external 158 dissolved oxygen field surrounding the instrument.

159 With a full sensor suite, the nominal power consumption of the PHORCYS is 50 mA at 12V; 160 standby current is 2 mA. When programmed to sample at a 50 % duty cycle, a single 12V primary 161 'D' cell battery pack (20,000 mAh capacity) will power the instrument for up to 30 d in an

162 unattended deployment mode. Data are recorded in an ASCII fixed-field format onto a micro SD card 163 in a DOS-readable format. For attended deployments, a combination communications and external 164 power port provides the ability to observe data in real time, allow program updates, download data, 165 and power the instrument indefinitely. The sampling interval is nominally set to 1 min, though data 166 can be collected as frequently as every 15 s. The acquisition program determines sampling activity by 167 way of a real-time clock. The chambers can thus be programmed to open and close at any time, 168 allowing the investigator to make multiple incubations of any desired length. In the configuration 169 used to acquire the data presented here, (Fig. 2, **D** symbol; e.g., Fig. 3) the chambers were 170 programmed to open at or around sunrise and the same operation was repeated at sunset, providing 171 two incubations in each 24 h period that aligned with the beginning and end of the photoperiod. The 172 chambers are opened and then closed sequentially (i.e., one after the other) to reduce total current 173 draw from the power source. The chambers remain open for 30 min at the outset of each incubation, 174 providing sufficient time for the water to be fully exchanged before closure; we confirmed this flushing time was sufficient in both quiescent and flowing (~ 1 m s^{-1}) waters using a series of tests 175 176 with a tracer dye (results not shown). While we used a standard 30 min flush time in the deployments 177 for which data is presented below, any time can be specified in the instrument's control software, 178 which is written in BASIC.

The earlier PHORCYS data from 2012 (Fig. 2, **O** symbols; e.g., Fig. 4) were obtained using a prototype instrument that permitted only one incubation cycle per deployment. This prototype (Fig. 1b) was assembled from two 2.5 L Niskin-style sampling bottles mounted to an aluminum frame (opaque polyvinylchloride and transparent polycarbonate plastic, respectively; actual usable volume, 2.6 L; General Oceanics, Inc., Miami, FL, USA). Closure of the chambers for incubation was effected using an electrolytic time release (i.e., "burn wire") system. Prior to deployment, the Niskin bottle endcaps were cocked open and the retaining cable was rigged to a fusible burn wire plug. Once in the water, a sufficient current was applied to the burn wire at a time set by the user, corroding the wire and allowing the bottle endcaps to close. The chambers were then sealed and the incubation began. For the deployments presented here, we programmed the chambers to close approximately 45 min after the PHORCYS had reached the desired depth. In the prototype instrument, Aanderaa model 4330F optodes (accuracy < 8 μ M O₂; resolution < 1 μ M O₂) were used to record dissolved oxygen concentrations.

192 Instrument deployments

193 We conducted 6 unattended deployments of the PHORCYS in 3 distinct ecosystem types in 194 the North Atlantic basin (Fig. 2; Table 2; Supplemental Table 1). Open-ocean deployments (2 to 7 d 195 in length, using the prototype instrument) were conducted during cruises aboard the R/V Knorr; 196 during these deployments, the instrument was suspended at various depths in the euphotic zone from 197 a drifting surface buoy (Fig. 1c). Deployment and recovery were accomplished in 45-60 min from a 198 standard oceanographic research platform (Supplemental Fig. 1). An Argo satellite beacon (Fig. 1c) 199 allowed us to track the array remotely between deployment and recovery while the ship traveled up 200 to 300 km away to conduct other shipboard scientific operations; we specifically designed both 201 models of the PHORCYS to be wholly autonomous, incurring a minimal burden on other shipboard 202 operations. Pierside deployments were conducted using the present, multiple-incubation version of the PHORCYS at the Iselin Marine Facility, Woods Hole, MA, USA (41° 31' 24" N 70° 40' 20" W); 203 204 the site adjoins a highly productive coastal embayment. In both cases, oxygen concentrations (umol L⁻¹ O₂), percent saturation, and temperature were then recorded for each chamber at 1 min intervals. 205 206 Post-acquisition corrections for salinity were applied to both the open-ocean and coastal data using 207 concurrent observations of salinity and manufacturer-supplied correction coefficients. Concurrent 208 salinity data were obtained from the continuous, flow-through CTD system aboard the R/V Knorr 209 (for open-ocean deployments) or from the Seabird CTD unit mounted as an external sensor on the

210 present PHORCYS model (2016 coastal deployments). Due to the limited number of at-sea 211 deployments of the PHORCYS prototype and the challenges we encountered during our initial 212 cruises, we pooled our prototype results with the data we later obtained from the production-model 213 instrument in order to assemble a larger, more robust dataset for subsequent analysis.

214 In

Instrument calibration and choice of deployment depth

215 Optodes were calibrated before each pierside deployment and prior to each research cruise 216 using a two-point method, assuming a linear response between end-members. An air-saturated 217 solution was obtained according to manufacturer instructions by bubbling ambient air for approx. 30 218 min through a sufficient volume of Milli-O water using an aquarium stone; a zero-oxygen solution 219 was obtained by dissolving an excess of reagent-grade sodium sulfite into a beaker containing Milli-220 Q water. The optodes were then calibrated at atmospheric temperature and pressure, as recommended 221 by the manufacturer. At the open-ocean stations, we deployed the prototype instrument within the 222 euphotic zone at depths where the observed flux of photosynthetically active radiation (PAR; 223 wavelengths 400-700 nm) was between 10-30% of the incident flux at the surface. The depth of the 224 euphotic zone (z_{eu} , defined as the depth at which PAR = 1 % of incident intensity) was identified 225 using profiles from shipboard hydrocasts. At each of these stations, we calculated an equivalent 226 deployment depth (z_{equiv}) that accounted for light attenuation by the transparent chamber's 227 polycarbonate plastic, according to a modification of the standard equation for the exponential decay 228 of light with depth:

229
$$I_z(PAR) = \frac{I_0(PAR)e^{-K_d(PAR)z_{equiv}}}{T(PAR)}$$
(1)

where $I_0(PAR)$ and $I_z(PAR)$ are, respectively, the incident PAR intensity and intensity at depth *z*, $K_d(PAR)$ is the diffuse attenuation coefficient for the PAR spectral band (calculated from the

hydrocast profile at each station), and T(PAR) is the transmissivity of the polycarbonate plastic expressed as a fraction (0.83). Pierside deployments of the present PHORCYS model were conducted at a depth of 1.5 m.

235 Instrument validation by two independent methods

First, to validate the optodes' ability to accurately track respiration, we used a standard 236 237 analytical method — two-point Winkler titration — to determine dissolved oxygen consumption in 238 triplicate water samples at the beginning and end of each incubation period (present model 239 instrument) or deployment (for data obtained with the prototype). Winkler titrations were conducted 240 in 125 mL BOD bottles according to EPA Method 360.2 as modified for shipboard determination in 241 seawater (Knapp et al., 1989). Initial Winkler titrations were made in samples collected within 15 242 min of deployment using a Niskin or Go-Flo bottle suspended at the same depth as the instrument. A 243 set of three darkened 125 mL BOD bottles containing water from the same Niskin or Go-Flo bottle 244 was incubated at in situ temperature until the PHORCYS was recovered or the incubation period 245 ended; these samples were then sacrificed according to the same protocol. The BOD bottles used in 246 these incubations were triple-rinsed with 10 % HCl and then Milli-Q water prior to sampling. All reagents for Winkler titrations were A.C.S. grade or better; the sodium thiosulfate titrant was 247 248 standardized daily. Amperometric titration was performed using an autotitrator (Metrohm 904 249 Titrando; Metrohm USA, Inc., Riverview, FL).

As an additional means of comparison during the open-ocean deployments, we also tracked changes in dissolved oxygen in a series of continuously monitored shipboard bottle incubations. Water from the PHORCYS deployment depth was retrieved for these incubations from a hydrocast made within 1 h of deployment. Incubations were conducted in gas-tight, 300 mL glass BOD bottles; at least 5 replicates were used for each series of measurements. Determination of dissolved oxygen

was made at 3 to 9 h intervals using optode spot minisensors (PreSens PSt3; response time < 40 s;
Precision Sensing GmbH, Regensburg, Germany; Warkentin et al., 2007) that were glued to the
inside surfaces of the bottles using food-quality silicone cement. The use of these optode spots
eliminated the need for drawing of aliquots from the sample bottles; the bottles had been soaked in
Milli-Q water for > 2 months following application of the sensor spots. Incubations were conducted
in the dark at *in situ* temperature as described in Edwards et al. (2011).

261 Data analysis

262 PHORCYS rate calculations

Volumetric rates of gross community respiration (GR) and net community production (NCP) were calculated by linear least-squares regression of observations of dissolved oxygen concentration over the length of the deployment (prototype model) or incubation period (present model). We assumed the only significant chemical reactions contributing to oxygen consumption and evolution in the two chambers were aerobic respiration and photosynthesis. For organic matter of elemental stoichiometry roughly corresponding to that of Redfield (1934) and assuming production is based on nitrate, these redox reactions can be represented as (after Stumm and Morgan, 2013):

270
$$1380_2 + C_{106}H_{263}O_{110}N_{16}P + 18HCO_3^- \rightarrow 124CO_2 + 16NO_3^- + HPO_4^{2-} + 140H_2O$$
 (2)

271 and

272
$$106CO_2 + 16NO_3^- + HPO_4^{2-} + 122H_2O + 18H^+ + h\nu \rightarrow C_{106}H_{263}O_{110}N_{16}P + 138O_2$$
 (3)

As is the case with the classic light and dark bottle incubation technique, we assumed that only respiration was taking place in the dark (i.e., GR; reported as positive quantities), while both reactions were taking place simultaneously in the presence of light (i.e. NCP; in reality, photorespiration also contributes to oxygen consumption in the light bottle). Finally, where possible,
we estimated the rate of gross primary productivity (GPP) according to the relationship

$$GPP = NCP + GR \qquad (4)$$

To estimate uncertainties in PHORCYS rate estimates, we adapted an effective degrees of freedom technique traditionally applied to time-series data in physical oceanography (Emery and Thomson, 2001). The method, which serves as an alternative to the standard error of the regression slope in instances when true replication cannot be achieved, is described in detail in the Appendix.

283 Rate calculations from Winkler titration samples and sensor spot incubations

For the Winkler titration samples, GR was calculated using a difference of means,

$$GR = -\frac{dO_2}{dt} \quad (5)$$

where dO_2 is the difference between the mean dissolved oxygen concentration in samples sacrificed 286 287 at the final incubation timepoint and the mean measured at the initial timepoint and dt is the 288 incubation duration. For the sensor spot incubations, we calculated rates for each bottle using a linear 289 least-squares regression of all observations recorded over the time period; we then averaged the rates 290 obtained for the various replicates to obtain a final estimate. Uncertainties in rates based on the 291 Winkler titration method were determined from the standard deviations of the dissolved oxygen 292 concentrations measured in the replicates at each timepoint. For rates based on the sensor spot 293 incubations, we used the standard error of regression. While we report metabolic rate estimates from 294 the PHORCYS in oxygen-based units, a stoichiometric relationship such as the molar respiratory 295 quotient of Anderson and Sarmiento (1994) can be applied to convert these estimates to units of 296 carbon.

297 Assessment

298 PHORCYS metabolic rate estimates

299 We observed a significant degree of daily and hourly variability in the time series data from 300 each PHORCYS deployment (e.g., Fig. 3 and 4). This variability manifested itself in a wide range of 301 metabolic rate estimates (Table 2; Supplemental Table 1) that reflect the interaction of multiple 302 biological and physical forcings, including diel changes in cellular growth cycle, surface-layer water 303 temperature, and irradiance. Daily rates of GR were estimated from dark chamber data for all 304 PHORCYS deployments (Table 2). For data obtained with the present model instrument, hourly rates 305 were calculated for each incubation segment (Fig. 3); these were extrapolated to daily rates to 306 facilitate comparison with other studies and with data from the prototype instrument. For open-ocean 307 data obtained with the PHORCYS prototype, daily rates were calculated using the entire DO time 308 series from each deployment. Erroneous readings from one of the optodes and a system malfunction 309 prevented us from recovering usable NCP data from the transparent chamber during two of the open-310 ocean deployments of the prototype instrument. During the 24-27 April 2012 deployment, the chosen 311 depth of 29 m provided insufficient PAR (< 3% of surface intensity) to support measurable 312 photosynthesis in the transparent chamber; consequently, we could not calculate NCP or GPP for this station (Supplemental Table 1). An obstruction prevented the transparent chamber from closing 313 314 during the November 2016 deployment, allowing us to recover useful data from only the dark 315 chamber.

We captured daily rates of GR ranging from $1.8 \pm 0.2 \ \mu \text{mol} \ O_2 \ L^{-1} \ d^{-1}$ at a mid-latitude station in the North Atlantic to 10.5 ± 7.5 and $18.9 \pm 1.9 \ \mu \text{mol} \ O_2 \ L^{-1} \ d^{-1}$ in two different water masses at the Woods Hole pier in early November (Fig. 3 and 4; Table 2). The wide variation in GR we observed with the PHORCYS covers a significant range of the rates for marine systems compiled by Robinson and Williams (2005). Daily rates of GPP at the PHORCYS deployment depth ranged from essentially zero on several days at a mid-latitude station in the North Atlantic (Supplemental Table 1) to $3.6 \pm 0.5 \mu mol O_2 L^{-1} d^{-1}$ during a coccolithophore bloom (Collins et al., 2015) in the sub-Arctic North Atlantic (Fig. 4). Rates of NCP at the deployment depth ranged from -2.0 ± 0.4 $\mu mol O_2 L^{-1} d^{-1}$ to $-4.2 \pm 0.2 \mu mol O_2 L^{-1} d^{-1}$ (Supplemental Table 1).

325 **Discussion**

326 Subdaily variation in rates of metabolism; choices concerning data analysis

327 Increasing evidence suggests the significant sub- and inter-daily variation in metabolic 328 activity captured by the PHORCYS exists in almost all natural aquatic systems (Caffrey, 2004; 329 Staehr et al., 2012; Collins et al., 2013); even in oligotrophic waters, respiration and production rates 330 may change significantly from one day (or hour) to the next, even as the system maintains an overall 331 state of near trophic balance (Aranguren-Gassis et al., 2012). The types of fluctuations we observed 332 in the various dissolved oxygen time series obtained from the PHORCYS appear to be characteristic 333 of incubation-based *in situ* instruments. McDonnell et al. (2015) and Boyd et al. (2015) both 334 observed similar patterns in dissolved oxygen data during recent deployments of an *in situ* device 335 that measures oxygen consumption rates on marine particles.

Nevertheless, the subtle changes in DO concentration we observed in multi-day incubations with the PHORCYS prototype (e.g., Fig. 4) suggest that a single regression line — fitted, in this case, to align with the closing and opening of the chambers — might have been poorly suited to a time series exhibiting such variation. While one might instead have divided the full time series into shorter segments to compute several different regressions (e.g., according to the photoperiod), we chose to define the interval for rate calculations from the prototype instrument according the chamber opening and closing times. This allowed us to avoid the bias inherent in dividing such a time series into 343 smaller segments. The present PHORCYS model allows users to predefine multiple, shorter 344 incubation periods of length appropriate to the ecosystem (e.g., Fig. 3); this feature eliminates the 345 need for a choice between either a subjective, *ex post facto* division of data or the application of a 346 single regression that might fail to capture the observed variation.

347 Evaluation of instrument performance using independent methods

348 PHORCYS estimates of community respiration were systematically higher than those 349 calculated for the same waters using the two-point Winkler titration method (Fig. 5a; Table 2). While there was a significant correlation between estimates from the two different methods ($r^2 = 0.42$; p =350 351 0.04), the traditional Winkler titration approach appeared to underestimate rates of respiration by 352 nearly one-third. Alternatively, the PHORCYS could have overestimated true rates of respiration 353 through artificial stimulation of the microbial community or inadvertent retention of residual 354 dissolved oxygen by the chambers or optode. In contrast, rate estimates from the PHORCYS 355 generally agreed with those based on our non-destructive optode sensor spot incubations, though we 356 had only five data points on which to evaluate the correlation (Fig. 5b; correlation was not 357 statistically significant).

358 Wikner et al. (2013) observed close agreement (~ 3 % deviation) between gross community 359 respiration rates derived from continuous optode measurements in a shipboard incubation chamber (1 360 L, clear glass) and a series of parallel incubations based on a Winkler method in 120 mL glass BOD 361 bottles. In previous work, we observed similar coherence between rates derived from a two-point 362 Winkler method in 300 mL glass BOD bottles and those based on incubations in 125 mL glass BOD 363 bottles fitted with optode sensor spots (surface area : volume ratios as reported in Table 1; Collins et 364 al., 2015). Underlying the agreement among methods in these previous studies is a common reliance 365 on incubations aboard ship. One might therefore conclude that the observed divergence between

estimates of respiration from the PHORCYS and those we made with the Winkler titration method
are due to inherent differences between the *in situ* PHORCYS approach and shipboard incubation
methods. However, evidence suggests that use of the same method is not even a guarantor of
agreement. For example, Robinson et al. (2009) found that vastly different rates were obtained from
the same method of measuring primary production when the only the timescale of incubation was
varied.

372 Possible sources of observed discrepancies between methods

373 *Gas permeability of materials*

374 The materials chosen for construction of the PHORCYS, particularly the polycarbonate 375 plastic we used for the incubation chambers, represent one possible source of bias in our method. We 376 chose polycarbonate material for its durability, low cost, and, compared with other plastics such as 377 polyvinylchloride, minimal biological reactivity. Polycarbonate was selected for its minimal 378 biological reactivity in construction of at least two similar in situ incubation devices (Langdon et al., 379 1995; Robert, 2012). However, polycarbonate, like most plastics, is at least partially permeable to 380 dissolved oxygen; by comparison, the borosilicate glass of which BOD bottles are fabricated is 381 nearly impermeable (Kjeldsen, 1993; Robert, 2012). While some diffusion of dissolved oxygen 382 across the polycarbonate chamber walls could have occurred during incubations with the PHORCYS, 383 thus biasing our results, we believe the effect would likely have been very modest given the 384 dissolved oxygen gradients and timescales typical of our deployments. First, the gradient necessary 385 to drive any diffusion across the PHORCYS chamber wall (i.e., the difference between the internal and external dissolved oxygen concentrations) did not exceed 25 μ umol O₂ L⁻¹ during any 386 387 deployment; in the deployment presented in Fig. 3, the average difference between internal and external concentrations was just 6.1 μ mol O₂ L⁻¹. In a series of oxygen-diffusion experiments with a 388

389 device similar to the PHORCYS, Robert (2012) monitored the dissolved oxygen concentration inside 390 a polycarbonate chamber after the water inside was treated with sodium hydrosulfite to render it 391 anoxic. After manipulating the dissolved oxygen concentration in water outside the chamber to create cross-boundary gradients ranging from approx. 50 μ mol O₂ L⁻¹ to > 350 μ mol O₂ L⁻¹, the authors 392 393 observed little change in the dissolved oxygen concentration inside the chamber on timescales similar 394 to those of the PHORCYS deployments (11-12 h for the current version of the instrument; 3-5 d for 395 the PHORCYS prototype). Further, in a comprehensive study of the permeabilities of several gases 396 (O₂, N₂, CO₂, and CH₄) in various polymers, Kjeldsen (1993) concluded that bias introduced by 397 permeability should be taken into account primarily when certain materials such as silicone rubber 398 were considered for use in anoxic waters; silicone rubber is more than 250 times as permeable to 399 dissolved oxygen than the polycarbonate plastic we used in construction of the PHORCYS (Kjeldsen, 400 1993).

401 Handling or sampling bias

402 Errors or bias introduced during the handling and manipulation of samples for bottle 403 incubations might also explain some of the discrepancy in rate measurements. For example, the 404 PHORCYS minimizes physical disturbances associated with seawater handling: Since the instrument 405 takes seawater samples and then incubates them in place, the planktonic community is not subjected 406 to rapid changes in temperature, pressure, and light associated with bringing water samples to the 407 surface via hydrocast and preparing them for shipboard incubations (Calvo-Díaz et al., 2011). The 408 PHORCYS also minimizes another potential bias that can arise when water containing marine 409 microbes is sampled from Niskin bottles; Suter et al. (2016) found that variation in settling rates 410 among marine particles can lead to an undersampling of microbial communities on faster-sinking 411 particles, which can fall below the bottles' spouts before aliquots can be drawn.

412 Biofouling and surface colonization

413 While we did not observe any significant biofouling of the PHORCYS or its components 414 during the deployments for which data are presented here, unwanted biological growth represents a 415 significant challenge and source of potential bias when autonomous sensors are deployed in the 416 surface and mesopelagic ocean (Delauney et al., 2010; Manov et al., 2004; e.g., Collins et al., 2013). 417 We imagine biofouling could represent a significant obstacle during future deployments if the 418 deployment duration were to exceed 5-7 d (the maximum explored in this work; Table 2), or if the 419 instrument were deployed in a productive ecosystem during a period of elevated primary production. 420 Compare, however, Robert (2012), who reported little biofilm growth on the chamber of a similar 421 incubation-based instrument after deploying it with no fouling controls for several months in the 422 Mediterranean Sea.

423 The lack of visual evidence of fouling in the PHORCYS chambers notwithstanding, there is 424 substantial evidence that microbial colonization of surfaces happens rapidly in the marine 425 environment (Salta et al., 2013; Dang and Lovell, 2016). It is therefore almost certain that at least 426 some microbial colonization of the chamber walls would have taken place within the 3-5 d timescale 427 of our deployments. Because we did not make measurements of biofilm activity or species 428 abundance, it is impossible to diagnose what contribution these communities might have made to the 429 observed fluxes in dissolved oxygen. However, compared to the surface area to volume ratios of the 430 150 and 300 mL vessels in which the other rate measurements were made (125 mL bottle, 0.83; 300 431 mL bottle, 0.77), the lower surface area to volume ratios of the PHORCYS chambers (current model, 432 0.34; prototype, 0.67) provide fewer opportunities for colonization relative to the volume being 433 incubated (Table 1). There is also some evidence that the small volumes typical of BOD bottles may 434 induce non-representative changes in the planktonic microbial community during incubations on 435 timescales of 24-48 h (Pratt and Berkson, 1959); however, compare Fogg and Calvario-Martinez

436 (1989), who found that such bottle size effects were only significant during periods of very high

437 primary productivity. A thorough, systematic assessment of these so-called wall and bottle effects

438 should be part of the community intercalibration we advocate in our conclusion.

439 Other possible sources of bias

440 Alternatively, the well-documented dependence of community respiration rates on 441 temperature (Yvon-Durocher et al., 2012) might explain the apparent disagreement between methods. 442 While temperatures within the PHORCYS chambers fluctuated only according to the movement of 443 tidal water masses (Fig. 3) or the natural warming and cooling of the surface layer (Fig. 4), the 444 temperature inside the incubator in which the Winkler samples were kept during the shipboard 445 deployments fluctuated during each series of experiments by $\pm 2^{\circ}$ C from the target. A further, more 446 intriguing explanation for the systematic discrepancy — pointing in this case to overestimation of 447 rates by the PHORCYS — is that dissolved oxygen could have accumulated on the optodes or inside 448 walls of the instrument chambers over the course of each deployment, in spite of the standard 30 min 449 flushing protocol. Wikner et al. (2013) offer convincing evidence for the accumulation of dissolved 450 oxygen on the optode surface and acrylic (polymethylmethacrylate) stopper used in their shipboard 451 incubation apparatus; the application of a correction factor for this bias reduced apparent rates 452 calculated from the optode time series relative to the Winkler method. We did not evaluate oxygen 453 retention by the polycarbonate plastic material of which the PHORCYS chambers were constructed.

The nature of linear regression itself may also play a role in differences observed between the PHORCYS and the Winkler-based method. A least-squares regression line fit to a large set of observations collected at high temporal frequency, such as those obtained from the PHORCYS, is sensitive in some degree to each of those observations. In comparison, a rate calculated from the beginning and ending oxygen concentrations in the two-point Winkler method is necessarily sensitive 459 only to those two observations; if some unrepresentative source of variability is present in just one 460 bottle, the entire measurement can be heavily biased. Regression of data from the PHORCYS thus 461 yields rate estimates which are robust to the sampling bias inherent in point measurements, yet the 462 technique leaves the underlying data intact to be further interrogated for information about natural 463 variability.

464 The "deep breath" phenomenon

465 Finally, while not evident in data from the deployments presented in Fig. 3 or 4, we did 466 occasionally observe a sharp initial decrease in dissolved oxygen concentration within the chambers 467 shortly after closure (data not shown). This phenomenon was reported extensively by Robert (2012) 468 during testing of a similar instrument in the Mediterranean Sea, and thus warrants further 469 investigation. The initial "deep breath" observed in these instances could reflect the rapid and 470 preferential utilization by the microbial community of a limited but highly labile fraction of the 471 dissolved organic carbon (DOC) pool within the chamber. Such a phenomenon has been observed 472 frequently in rates of DOC consumption in freshwater systems, where labile carbon is often 473 metabolized at a rapid rate in the initial minutes or hours of an incubation, leading to an apparent 474 decline in metabolic activity once the labile pool has been exhausted and the incubation progresses 475 (del Giorgio and Pace, 2008; Guillemette and del Giorgio, 2011; Guillemette, McCallister, et al., 476 2013). While we did not directly observe a pronounced initial change in the rate of dissolved oxygen 477 consumption in any of our shipboard optode sensor spot incubations, it is possible that our sampling 478 interval (3 to 9 h; see above) was of insufficient resolution to detect it.

479 Comments and recommendations

480 Through autonomous collection of biogeochemical observations at uniquely high temporal
481 frequency, the PHORCYS yields estimates of community metabolic activity while simultaneously

482 freeing the analyst from the logistical constraints of attended water column sampling and preparation 483 of shipboard incubations. While we could not determine the origin of the systematic discrepancy 484 between the PHORCYS rate estimates and those based on the traditional two-point Winkler method, 485 the instrument's design allows investigators to avoid many of the potential biases associated with 486 bottle incubations that have been previously documented in the literature. The PHORCYS offers a 487 further advantage in that it can be used to collect information over multiple timescales about the 488 metabolic state of marine and aquatic ecosystems at minimal cost and burden to the user. While the 489 mixed-layer deployments we have presented here provided volumetric rates of ecosystem metabolism 490 at a single depth, multiple PHORCYS units could be deployed simultaneously at different depths in 491 the water column as a means of making depth-integrated rate measurements; one could compare 492 these rates to estimates obtained from *in situ* geochemical tracer studies. Future work could also 493 include direct, side-by-side comparison of metabolic rates obtained from the PHORCYS with those from the classic, bottle-based ¹⁴C or ¹³C tracer methods. 494

495 In this spirit, we believe a new, community intercalibration effort is warranted to 496 systematically evaluate the many sources of uncertainty in incubation-based measurements of 497 community metabolism. The workshops of the Group for Aquatic Primary Productivity (GAP; 498 Figueroa et al., 2014) could serve as a model for such an effort, which should include comprehensive 499 evaluation of various combinations of incubation chamber materials, types of oxygen sensors, 500 chamber sizes, and incubation durations. Genetic and biogeochemical tools for characterizing the 501 extent, mechanisms, and effects of surface colonization within the incubation chambers would be 502 critical to the success of any such endeavor. We are optimistic such an effort might reveal the causes 503 of longstanding discrepancies such as those we observed between our PHORCYS rate estimates and 504 those we obtained with bottle-based methods.

505 Acknowledgments

506 We thank the captains and crews of the R/V Knorr and R/V Clifford Barnes, Anton Zafereo, 507 Kay Bidle, Bethanie Edwards, Filipa Carvalho, Jared Schwartz, Fiona Hopewell, Gabriel Roy 508 Liguori, Richard Payne, Jason C. Smith, Sujata Murthy, Dave Fisichella, Ed O'Brien, Craig 509 Marquette, Erik Smith, Shawn Sneddon, Richard Butler, Helen Fredricks, David Glover, Oliver 510 Newman, Emily Peacock, Leah Houghton, Matthew Bogard, Olivia De Meo, and Joe Salisbury. 511 Sheri White contributed significantly to early development of the PHORCYS. The comments of two 512 reviewers significantly improved our original manuscript. This research was supported by the U.S. 513 National Science Foundation (awards OCE-1155438 to B.A.S.V.M., J.R.V., and R.G.K., and OCE-514 1059884 to B.A.S.V.M.), the Woods Hole Oceanographic Institution through a Cecil and Ida Green 515 Foundation Innovative Technology Award and an Interdisciplinary Science Award, and a U.S. 516 Environmental Protection Agency (EPA) STAR Graduate Fellowship to J.R.C. under Fellowship 517 Assistance Agreement no. FP-91744301-0. The publication has not been formally reviewed by EPA. 518 The views expressed in this publication are solely those of the authors, and EPA does not endorse 519 any products or commercial services mentioned in this publication. 520 Availability of data and code 521 A MATLAB script to read in, process, and estimate rates and uncertainties in dissolved

522 oxygen data from the PHORCYS is provided online at

523 <u>https://github.com/jamesrco/DO_Instruments/</u>. The script can be easily adapted to calculate N_{eff} -

- 524 based estimates of uncertainty in any dissolved oxygen time series. All PHORCYS and Winkler
- 525 titration data and other scripts required to reproduce the results and figures in this work are available
- 526 online in the same location.

528 Appendix

529 A new method for calculation of uncertainties in metabolic rate estimates

530 The ideal means of estimating uncertainties in PHORCYS rates would have been true 531 biological replication, i.e., the simultaneous deployment of several identical instruments in the same 532 water mass. One could then have used the standard deviation of the rate measurements in each 533 different instrument as an estimate of the overall uncertainty. Because we had only one instrument — 534 an exceedingly common situation in oceanographic work — such true replication was not possible. 535 The standard error of the regression slope provides one possible estimate of uncertainty in time-series 536 dissolved oxygen data; for example, this common approach was recently applied to data from *in situ* 537 chamber incubations of sinking marine particle material (McDonnell et al., 2015). However, we 538 assumed that the standard error of regression would significantly underestimate the true uncertainty 539 in our estimates since it does take into account the reduced number of degrees of freedom in such a 540 time series. Because the data points in such a dissolved oxygen time series are not independent of one another, there are almost always far fewer effective degrees of freedom N_{eff} in such data than the 541 542 number of observations (i.e., data points) N (Emery and Thomson, 2001). (We represent the effective degrees of freedom by the notation N_{eff} in lieu of the N^* notation used by Emery and Thompson.) 543

544 Our approach was the following: For each time series of dissolved oxygen concentrations, we 545 first approximated the integral time scale *T* of the data according to

546
$$T = \frac{1}{C(0)} \int_0^{B_0} C(\tau) d\tau$$
 (A1)

where C(0) is the value of the autocorrelation function *C* of the time series at lag $\tau = 0$, and B_0 is the time lag of the autocorrelation function at the first zero crossing of the *x*-axis, which (after, e.g., Talley et al., 2011) we use as an estimate of the timescale of decorrelation. We then followed the 550 method of Emery and Thompson (2001) to estimate the effective number of degrees of freedom N_{eff} 551 from *T*, *N*, and Δt , where Δt is the sampling interval of the data:

552
$$N_{eff} = \frac{N\Delta t}{T} \quad (A2)$$

In this formulation, $N\Delta t$ is therefore the total length of the oxygen time series in which the rate estimate was made. Finally, we used this N_{eff} to obtain $s_{\hat{\beta}_{1,adj}}$, an adjusted estimate of the standardized uncertainty in the slope parameter of the regression (i.e., $\hat{\beta}_1$, the rate of dissolved oxygen consumption or production), according to

557
$$s_{\hat{\beta}_{1,adj}} = \sqrt{\frac{SSE}{N_{eff}-2}\frac{N}{\Delta}} \quad (A3)$$

where *SSE* is the sum of the squared errors from the fit of the regression, i.e., $\sum_{i=1}^{N} (y_i - \hat{y}_i)^2$, *N* is the number of observations (as above), and Δ is the determinant $NS_{xx} - S_x^2$. Eq. A3 is simply the formula for calculation of the standard error of the regression slope in the unweighted case, except that N_{eff} is used instead of *N*.

This method of estimating uncertainties in PHORCYS rates produced values of N_{eff} , the 562 563 number of effective degrees of freedom, which were typically $\ll N$, the number of observations in the given dissolved oxygen time series (Table A1). Estimates of the integral time scale T ranged from 564 565 0.5 to 7.2 h; at station PS-2, the 77.4 h deployment for which data are presented in Fig. 4, we 566 estimated T to be 7.2 h (Table A1). Using the N_{eff} derived from these time scales, we obtained adjusted uncertainty estimates for our PHORCYS rates $(s_{\hat{\beta}_{1,adj}})$ which were much greater in each 567 case than the standard error of the regression slope, $s_{\hat{\beta}_1}$ (compare mean precision of 24.8 % and 3.4 568 %, respectively; Table A1). While more robust than the corresponding $s_{\hat{\beta}_1}$, these $s_{\hat{\beta}_{1,adj}}$ still reflect a 569

- 570 fundamental limitation of linear regression: Both methods yield estimates of uncertainty which are
- 571 inversely proportional to the number of data points (i.e., the length of the underlying data series) and
- 572 the range of values spanned by the independent variable.

573 Figures







Fig. 2. Locations of PHORCYS deployments described in the text. Primary map: Unattended openocean deployments from a surface mooring were conducted using the PHORCYS prototype at 5 stations during two cruises aboard the *R/V Knorr*. Stations QL-1 and QL-2 were conducted during cruise KN207-1; PS-1, PS-2, and PS-3 were conducted during cruise KN207-3. Inset: Pierside deployments using the present PHORCYS model were conducted in November 2016 at the Iselin Marine Facility in Woods Hole, MA, USA.



591

592 Fig. 3. Continuous, unattended observations of community respiration at the Iselin Pier in Woods 593 Hole over a 3 d period in November 2016. (a) Record of dissolved oxygen concentration in the 594 opaque PHORCYS chamber. The incubation periods from which estimates of GR were calculated are 595 separated by 30 min, *in situ* flushing periods when the chamber was opened and closed to obtain a new water sample. Separate estimates of GR for each of the incubation periods (units of μ mol O₂ L⁻¹ 596 d^{-1}) were obtained by linear least-squares regression, and are shown as solid traces superimposed 597 598 over the instrument data; these are the Iselin Pier data reported in Tables 2 and A1. Uncertainties 599 were determined using the effective degrees of freedom method described in the Appendix. (b) 600 Ambient *in situ* dissolved oxygen concentration measured concurrently outside the PHORCYS 601 chambers, reflecting tidal changes in water-mass properties. (c) Ambient in situ temperature data

- 602 recorded outside the PHORCYS chambers. An instrument malfunction during the deployment
- 603 prevented us from recovering data from the transparent PHORCYS chamber.



605 Fig. 4. Unattended observations of ecosystem metabolism made with the prototype PHORCYS 606 instrument at station PS-2 (Fig. 2; Supplemental Table 1) during a sub-Arctic, open-ocean 607 deployment aboard the R/V Knorr. A midsummer bloom of a calcifying phytoplankton species was 608 in progress at the site (Collins et al., 2015) when these observations were made. (a) Estimates (in units of umol $O_2 L^{-1} d^{-1}$) of gross community respiration (GR) and net community production (NCP) 609 610 were obtained by linear least-squares regression, and are shown as traces (GR as solid trace; NCP as 611 dashed trace) superimposed over the instrument data. The prototype instrument allowed for only a 612 single incubation over the course of the deployment. GPP was calculated as the difference between 613 GR and NCP based on Eq. 4 in the text. Uncertainties were determined using the effective degrees of 614 freedom method described in the Appendix. Incident photosynthetically active radiation (PAR) was 615 measured using shipboard sensors. Dissolved oxygen concentrations measured concurrently in dark 616 shipboard incubations using optode sensor spots (5 replicates; error bars show \pm SD) are 617 superimposed as open circles. The respiration rate estimated from these incubations is shown as a 618 dotted trace. (b) Diel warming of the surface layer is evident in *in situ* temperature data collected by 619 the PHORCYS.



Fig. 5. Comparison of community respiration (GR) rate estimates from the PHORCYS (*x*-axis) with rates determined by (a) the two-point Winkler titration method and (b) a series of shipboard bottle incubations using optode sensor spots. Circles show data from the prototype instrument, while squares show data collected with the present PHORCYS model. A Type II (major axis) regression (dashed line) was fit to each set of paired observations using the lmodel2 package for R (Legendre, 2014). In (a), the regression model was fit to a single dataset consisting of both current and prototype model data. A red 1:1 line is superimposed in each panel for reference.

628

629 **Table 1.** Estimated surface area to volume ratios of PHORCYS chambers and standard BOD bottles.

630

| Bottle or chamber type | Actual usable | Estimated | Estimated surface |
|---------------------------------|-----------------|-------------------------|---------------------|
| | volume (mL) | internal surface | area : volume ratio |
| | | area (cm ²) | |
| PHORCYS chamber (prototype) | 2610 | 1760 | 0.67 |
| PHORCYS chamber (current model) | 5680 | 2035 | 0.34 |
| Typical 125 mL BOD bottle | 149.2 ± 0.3 | 124.4 ± 5.0 | 0.83 |
| Typical 300 mL BOD bottle | 299.2 ± 0.4 | 229.1 ± 4.3 | 0.77 |

631

632 The average volumes and surface areas reported in this table for BOD bottles were determined from

633 independent measurements of the dimensions of 10 different bottles of each size; these were chosen

at random from the Woods Hole Oceanographic Institution inventory.

| Deployment dates | Incubation period | Incubation duration (h) | Location ^a | PHORCYS model ^b | Community respiration (GR) (μ mol O ₂ L ⁻¹ d ⁻¹ ± uncertainty) | | |
|---------------------|-------------------|-------------------------------|-----------------------|-------------------------------|---|---------------------------------------|---|
| | | | | | PHORCYS opaque bottle ^c | Shipboard incubations ^d | Two-point difference of Winkler titrations at $t=0$ and recovery ^e |
| 24-27 Apr 2012 | Entire deployment | 71.6 | QL-1 | Prototype | 1.8 ± 0.2 | 3.2 ± 0.7 | 0.6 ± 0.1 |
| 30 Apr - 3 May 2012 | Entire deployment | 65.4 | QL-2 | Prototype | 4.2 ± 0.3 | 1.1 ± 0.2 | 1.2 ± 0.04 |
| 17-19 June 2012 | Entire deployment | 41.2 | PS-1 | Prototype | 2.4 ± 0.3 | 3.4 ± 0.5 | 3.5 ± 0.2 |
| 23-27 June 2012 | Entire deployment | 77.4 | PS-2 | Prototype | 7.8 ± 0.4 | 4.0 ± 0.3 | _ |
| 7-11 July 2012 | Entire deployment | 94.0 | PS-4 | Prototype | 6.0 ± 0.5 | 7.9 ± 0.6 | 7.4 ± 0.2 |
| 7-8 Nov 2016 | 17:15-06:00 | 12.7 | Iselin Pier | Present model | 18.9 ± 1.9 | | 5.9 ± 1.0 |
| 8 Nov 2016 | 06:15-16:45 | 10.5 | Iselin Pier | Present model | 2.2 ± 1.6 | | -0.8 ± 2.4 |
| 8-9 Nov 2016 | 17:20-06:00 | 12.7 | Iselin Pier | Present model | 8.0 ± 1.9 | _ | 4.3 ± 0.9 |
| 9-10 Nov 2016 | 17:30-06:00 | 12.5 | Iselin Pier | Present model | 10.5 ± 7.5 | | 5.4 ± 1.3 |

Table 2. Rates of community respiration measured in opaque bottles using the PHORCYS and two independent, traditional methods.

^a Cruise station or geographical location (Fig. 2); additional metadata for each station are provided in Supplemental Table 1

^a See Fig. 1

^c Uncertainty adjusted for effective degrees of freedom, as described in the Appendix ^d Mean of \geq 5 replicates; uncertainty derived from standard error of regression slope

^e Mean of 3 replicates; uncertainty derived from standard error

Table A1. Comparison of methods for estimation of uncertainties in dissolved oxygen time series data.

643

| Deployment dates | Location ^a | PHORCYS community respiration (GR) (µmol O ₂ | No. obser- vations (N) | Incubation duration (h) | Est. integral time scale (h) | Effective degrees of freedom (N_{eff}) | Estimated uncertainty $(\mu mol O_2 L^{-1} d^{-1})$ | | Method precision (est. uncertainty as percent of rate measurement) | |
|---------------------|-----------------------|---|---------------------------------|-------------------------------|--|--|---|------------------------------------|--|-------------------------------|
| | | $L^{-1} d^{-1}$) | | | | | Standard error of | Adjusted estimate | $S_{\hat{\beta}_1}$ | $S_{\widehat{\beta}_{1,adj}}$ |
| | | | | | | | regression slope $(s_{\hat{\beta}_1})$ | $S_{\hat{\beta}_{1,adj}}$ based on | | |
| | | | | | | | | N _{eff} | | |
| 24-27 Apr 2012 | QL-1 | 1.8 | 2150 | 71.6 | 1.1 | 66.5 | 0.03 | 0.18 | 1.6 % | 9.9 % |
| 30 Apr - 3 May 2012 | QL-2 | 4.2 | 1964 | 65.4 | 3.1 | 21.0 | 0.03 | 0.28 | 0.7 % | 6.7 % |
| 17-19 June 2012 | PS-1 | 2.4 | 1238 | 41.2 | 0.5 | 76.4 | 0.08 | 0.32 | 3.3 % | 13.1 % |
| 23-27 June 2012 | PS-2 | 7.8 | 2323 | 77.4 | 7.2 | 10.7 | 0.03 | 0.43 | 0.4 % | 5.5 % |
| 7-11 July 2012 | PS-4 | 6.0 | 2820 | 94.0 | 1.9 | 49.8 | 0.07 | 0.52 | 1.2 % | 8.6 % |
| 7-8 Nov 2016 | Iselin Pier | 18.9 | 765 | 12.7 | 1.3 | 19.8 | 0.29 | 1.87 | 1.5 % | 9.9 % |
| 8 Nov 2016 | Iselin Pier | 2.2 | 627 | 10.5 | 1.2 | 17.0 | 0.25 | 1.60 | 11.6 % | 74.4 % |
| 8-9 Nov 2016 | Iselin Pier | 8.0 | 760 | 12.7 | 1.6 | 16.2 | 0.26 | 1.89 | 3.2 % | 23.6 % |
| 9-10 Nov 2016 | Iselin Pier | 10.5 | 750 | 12.5 | 2.9 | 8.7 | 0.71 | 7.51 | 6.7 % | 71.4 % |
| Mean | | | | | 2.3 | | | | 3.4 % | 24.8 % |

^a Cruise station or geographical location (Fig. 2); additional metadata for each station are provided in Supplemental Table 1

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